

A study on Adsorption of Copper (II) Ions in Aqueous Solution by Chitosan - Cellulose Beads Cross Linked by Formaldehyde

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ABSTRACT

The effective removal of heavy metals from industrial effluents is one of the vital issues throughout the world. Removal of Cu^{+2} ions in aqueous solution have been analysed by using Chitosan beads made by blending with Cellulose and cross linking by Formaldehyde. Batch adsorption experiments were carried as a function of adsorbent dosage, pH, contact time, initial metal ion concentration and temperature. The optimum pH was found to be 5. The experimental data were tested with Langmuir, Freundlich, Temkin and Dubinin-Radushkevich isotherms and the data have been fitted very well with the Freundlich isotherm. The energy of adsorption showed that the adsorption of Copper by Chitosan - Cellulose beads was physical adsorption. Adsorption kinetics data were modeled with the application of Pseudo first order, Pseudo second order, Elovich and Intra-particle diffusion models. The results revealed that the Pseudo second order model was the best fitting model. The adsorption mechanism followed two stages in which the first one was rapid and the other was slower. The Boyd plot exposed that the intra-particle diffusion was the rate controlling step of the adsorption of Copper (II) ions by Chitosan-Cellulose beads.

keywords: *Heavy metal removal, Chitosan-Cellulose beads, Adsorption isotherms, Kinetics.*

1. INTRODUCTION

Heavy metal contamination of various water resources is a crucial concern globally due to their toxicity to human and other living beings. Industries such as Chemical, Leather, Mechanical and Electrical are the significant sources of heavy metal pollution. Copper is an essential trace nutrient that is required in small amounts (1-1.5 mg per day in food) by humans, other mammals, fish and shell fish for the synthesis of haemoglobin, carbohydrate metabolism and the functioning of more than 30 enzymes.

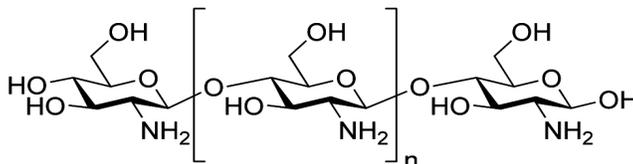
Although Copper can be an essential trace element, it could be harmful when it exceeds the tolerance limit. Copper is extensively used in the manufacture of fungicides, anti-fouling paints and also in electrical industries. Copper fume causes irritation of the eyes, nose and throat, headaches, stomach aches, dizziness, vomiting and diarrhoea and an illness called metal fume fever. High uptakes of copper may cause liver and kidney damage and even death. When copper ends up in soil, it strongly attaches to organic matter and minerals. As a result it does not travel very far after release and it hardly ever enters groundwater. In surface water, Copper can travel great

distances, either suspended on sludge particles or as free ions. Copper does not break down in the environment and because of that it can accumulate in plants and animals when it is found in soils. On Copper-rich soils, only a limited number of plants have the chance of survival and hence there is not much plant diversity near copper disposing factories.

Heavy metal ions are non-biodegradable and they have to be removed from water sources by various physical and chemical methods like chemical precipitation, evaporation, electrolysis, ion exchange, membrane separation and adsorption¹. In particular, adsorption is an effective and economic method for removal of pollutants from waste water². Many materials of biological origin have been utilised as adsorbents to remove heavy metal ions from water and industrial effluents.

Chitosan has been recognised as a biopolymer with significant potential for use as biosorbent for removal of metal ions from wastewater. Chitosan is commercially produced by the deacetylation of Chitin which is found in the outer skeleton of shrimp, crab, lobster and crayfish shells. Chitosan is a linear

polysaccharide composed of randomly distributed β -(1-4)-linked D-glucosamine (deacetylated unit) and N-acetyl-D-glucosamine (acetylated unit). Adsorption experiments with Copper³⁻⁶, Mercury⁷, Chromium^{8,9} and Lead¹⁰ ions indicated that Chitosan can be effectively used to adsorb these metals by establishing their different interactions with its amino and hydroxyl groups. Chitosan is the second most abundant bio polymer in earth after cellulose. Consequently,



Chitosan offers a lot of promising benefits for wastewater treatment applications today. Cost of Chitosan is much lower than activated carbon and it has excellent binding capacity¹¹. Chitosan has the characteristic feature of having amine groups in which nitrogen is a donor of electron pair that is attractive to most heavy metals and OH groups also take part in the adsorption¹².

However, unlike Chitin, Chitosan is soluble in acids. Although cross linking reduces adsorption capacity, it improves the resistance of Chitosan in both acid and alkaline medium. Hence the cross-linking widens the application of Chitosan in a broad pH range¹³.

In this research work, adsorption of Copper by Chitosan blended with Cellulose and cross linked by Formaldehyde has been analysed. The adsorbent has been made in the form of small beads of about 1.5 mm diameter. Cellulose is used as the blending polymer since it has similar structure as Chitosan and thus a homogeneous blend can be produced. Also, Cellulose is the most abundant natural polymer and Cellulose can be used as an adsorbent for its carboxyl and hydroxyl functional group which becomes the active binding site of the metal¹⁴⁻¹⁶. Hence the blending of Cellulose is expected to enhance the adsorption process.

2. MATERIALS AND METHODS

2.1 MATERIALS

Chitosan flakes with a deacetylation degree of 85 % were acquired from Pelican Biotech & Chemicals Labs Pvt. Ltd., Kuthiathode, Kerala, India. The chemicals used in this study such as Acetic acid, Cellulose powder, Sodium hydroxide pellets, Formaldehyde solution were in AR grade and manufactured by SD Fine Chem Limited, Mumbai, India. The AR grade of Cupric Sulphate penta hydrate ($\text{CuSO}_4 \cdot 5\text{H}_2\text{O}$) was used for the preparation of Cu^{+2} ions. Distilled water has only been used to make all solutions in this experimental study.

2.2 METHOD OF PREPARATION OF ADSORBENT (CHITOSAN-CELLULOSE BEADS)

2 g of Chitosan were dissolved in 200 ml of 1% Acetic acid and stirred for 3 hours to make a Chitosan gel. Then 2 g of Cellulose powder was added and stirred for 3 hours for uniform mixing. Formaldehyde solution (10% v/v) was added for the cross linking of polymer chain and stirring continued for another 3 hours. Then the Chitosan - Cellulose gel was injected through a syringe (without needle) over the surface of 1 M NaOH solution in a wide glass tray. The Chitosan - Cellulose beads were obtained on the surface of NaOH solution and they were allowed to stay in it for 12 hours. Then the beads were carefully separated from NaOH solution, cautiously washed many times with distilled water and allowed to be dried for 48 hours at room temperature.

2.3 ADSORPTION EXPERIMENTS

Adsorption of Cu^{+2} ions was carried out in batch process with initial concentration ranged from 100 ppm to 500 ppm. Cu^{+2} solutions of necessary concentrations were prepared by dissolving Cupric Sulphate penta hydrate ($\text{CuSO}_4 \cdot 5\text{H}_2\text{O}$) in distilled water. Batch adsorption experiments were carried out in 250 ml glass beakers filled with 100 ml of solution. Beads of adsorbent were added in the beaker and stirred by mechanical stirrer at 250 rpm. The concentration of Cu^{+2} ions after various adsorption processes were analysed by UV-Vis Spectrophotometer under visible lamp range with a wave length of 820 nm.

Equilibrium adsorption capacity (q_e) = $[(C_0 - C_e) / W] * V$

Where, C_0 and C_e are the initial and final Cu^{+2} concentrations (mg/L) of the solution in each adsorption experiment. V is the volume of the

Copper solution in litres, W is the weight of adsorbent in each beaker in grams and q_e is in mg/g.

Chitosan – Cellulose beads changed to blue colour after Cu^{+2} adsorption by simply showing that Copper ions were chelated.

3. RESULT AND DISCUSSIONS

3.1 EFFECT OF ADSORBENT DOSAGE

Adsorbent dosage has strongly affected the sorption capacity. With the fixed metal ions concentration, the percentage removal of metal ions increases with increasing weight of the adsorbents. This was due to more

availability of active sites or surface area at higher concentration of adsorbent.

Adsorption experiments of various dosages starting from 0.05 g to 0.2 g were carried out at room temperature (28°C) in separate 250 ml beakers and each beaker contained 100 ml of 100 ppm concentration. The pH of the solution was 5.2. The samples were tested in every 15 minutes time interval. Among them, 0.2 g of Chitosan - Cellulose beads were found effective and it derived 100 % adsorption of Copper in 100 ppm solution in 120 minutes.

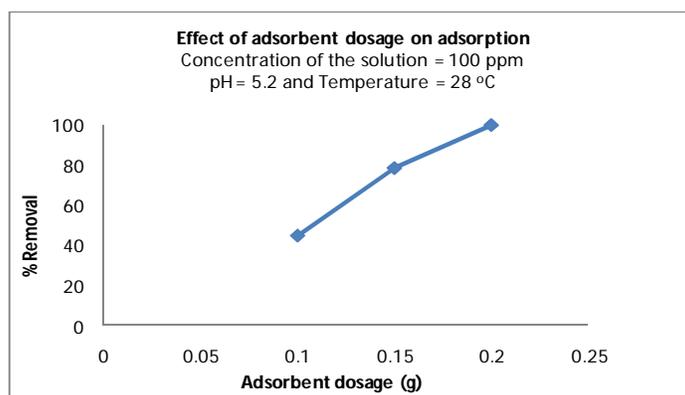


Fig.1: Effect of adsorbent dosage on the adsorption of Copper on to Chitosan-Cellulose beads

3.2 EFFECT OF pH

The batch experiments were carried out with a pH range of 2 to 10. A wide pH range has been used also to test the insolubility of the adsorbent in strong acidic and alkaline media. 1 M HCl and 1 M NaOH solutions were used to alter pH of the solution. The result showed that there was no adsorption at pH of 2. The adsorption of Copper reached maximum at pH of 5. The adsorption slowly decreased from pH of 6. The pH of the solution had significant impact on the uptake of heavy metal ions.

Hence it was clear that the adsorption of the adsorbent was pH dependent. According to Low et al., little sorption at lower pH could be ascribed to the hydrogen ions competing with metal ions for sorption sites¹⁷. At higher pH range, the Copper ions precipitated as their hydroxides which decreased the adsorption rate and as a result of reduction in the percentage removal of Copper ions. The cross-linked Chitosan – Cellulose beads proved a good chemical stability in the pH range of 2 to 10.

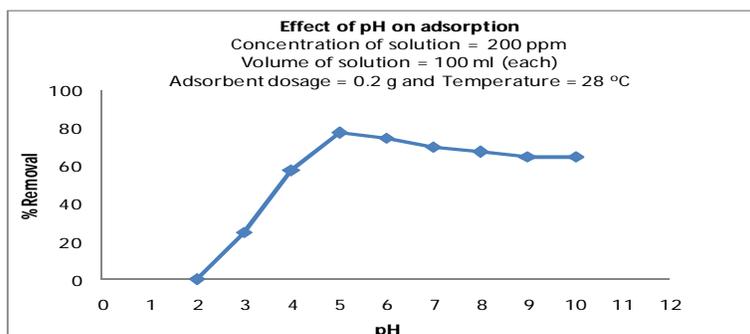


Fig.2: Effect of pH on the adsorption of Copper on to Chitosan-Cellulose beads

3.3 EFFECT OF CONTACT TIME ON ADSORPTION

The removal of Copper ions increased with time and attained saturation in about 180 minutes. The removal was very fast at the beginning and it gradually decreased with time till it reached equilibrium. The experimental data showed a considerably fast adsorption

during the first 15 minutes of adsorbent - adsorbate contact and it slowly decreased with time due to the saturation of the adsorption sites. Hence a two stage adsorption mechanism with the first rapid and the second slower had been seen.

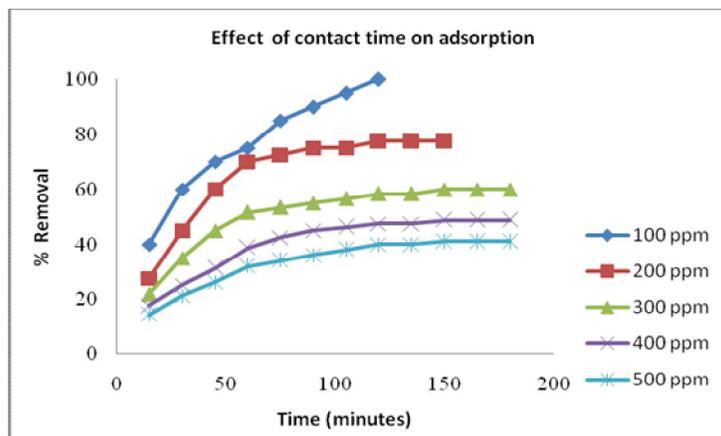


Fig. 3: Effect of Contact time on the adsorption of Copper on to Chitosan-Cellulose beads

3.4 EFFECT OF INITIAL METAL ION CONCENTRATION ON ADSORPTION

The metal uptake mechanism depended on the initial metal ion concentration. Metals were adsorbed by specific sites at low concentrations. But the adsorption amount did

not increase proportionally for higher metal ion concentrations since the active sites were filled and saturated. Hence it was very clear that the percentage removal of metal ion decreased with increase in metal ion concentration.

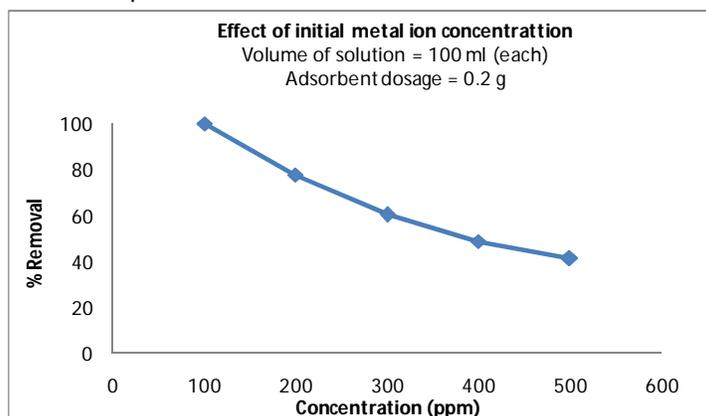


Fig. 4: Effect of initial metal ion concentration on the adsorption of Copper on to Chitosan-Cellulose beads

3.5 EFFECT OF TEMPERATURE

A temperature range started from 30° C in the multiples of 5° C were analysed up to 55° C. The adsorption increased slightly (2 %)

between 35° C and 45° C and then decreased up to 10 % when the temperature was further raised.

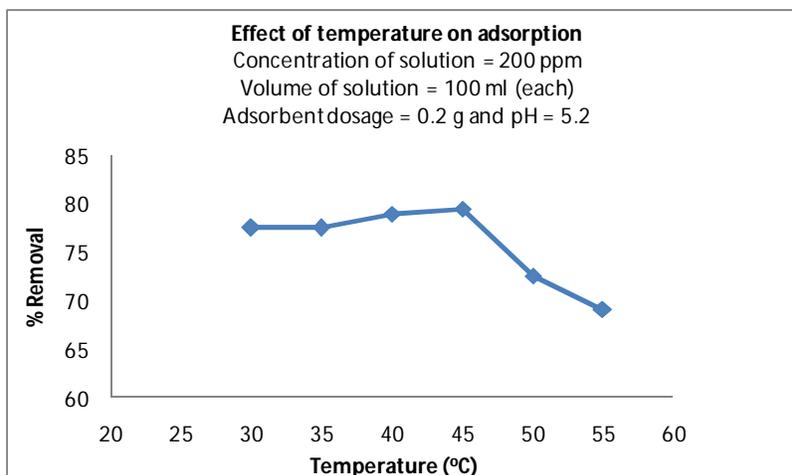


Fig. 5: Effect of temperature on the adsorption of Copper on to Chitosan-Cellulose beads

3.6 ADSORPTION ISOTHERMS

Adsorption isotherms describe the interaction of adsorbates with adsorbents. The experimental adsorption data of Copper (II) ions on the cross linked Chitosan - Cellulose beads were analysed by Langmuir, Freundlich, Temkin and Dubinin-Radushkevich isotherms. The experimental data showed that the adsorption of Copper (II) ions by the Chitosan-Cellulose beads increased with an increase in initial metal ion concentration significantly. At lower initial Copper ion concentrations, the adsorption increased linearly. At higher initial Copper ion concentrations, the adsorption capacity did not increase proportionally due to

the limitation of number of active sites on the surface of adsorbent beads.

3.6.1 LANGMUIR ISOTHERM

The Langmuir adsorption model is based on the assumption that maximum adsorption corresponds to a saturated monolayer of solute molecules on the adsorbent surface, with no lateral interaction between the adsorbed molecules. The Langmuir adsorption isotherm has been successfully used in many monolayer adsorption processes. The adsorption isotherm data were analysed by the Langmuir isotherm model in the linearised form,

$$C_e/q_e = C_e/q_{max} + 1/(b q_{max})$$

Where q_e is the equilibrium adsorption capacity of the adsorbent (mg/g), C_e is the equilibrium Cu concentration in solution (mg/l), q_{max} is the maximum amount of Cu that could be adsorbed on the adsorbent (mg/g) and K_L is the Langmuir adsorption equilibrium constant (L/mg). The plot of C_e/q_e versus C_e is shown in figure below.

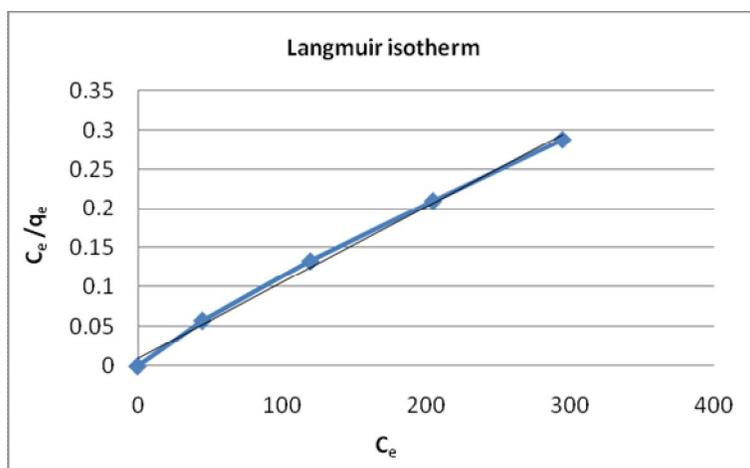


Fig. 6: Langmuir isotherm plot for the adsorption of Copper by Chitosan-Cellulose beads

3.6.2 FREUNDLICH ISOTHERM

The Freundlich model can be applied to multilayer adsorption with non-uniform distribution of adsorption heat and affinities over the heterogeneous surface. The experimental data were analysed by Freundlich isotherm model in the linearised form,

$$\log q_e = 1/n \log C_e + \log K_F$$

Where K_F is the Freundlich adsorption constant and it is the maximum adsorption capacity of metal ions (mg/g) and n is the constant illustrates the adsorption intensity (dimensionless). The plot of $\log q_e$ versus $\log C_e$ is shown below.

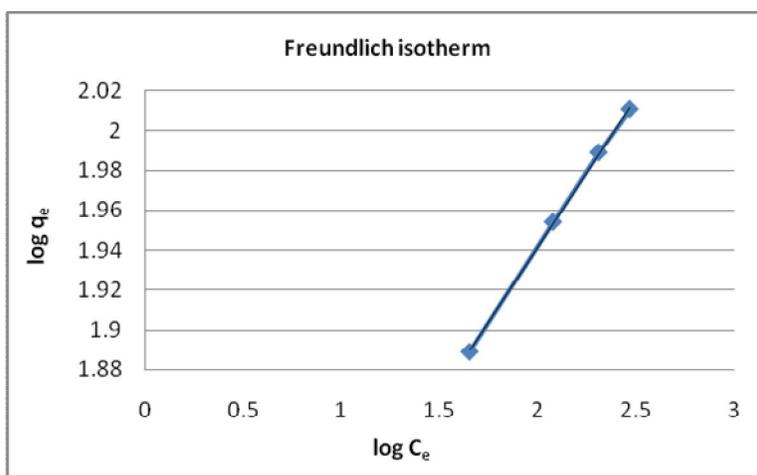


Fig. 7: Freundlich Isotherm plot for the adsorption of Copper by Chitosan-Cellulose beads

3.6.3 TEMKIN ISOTHERM

The derivation of the Temkin isotherm assumes that the fall in the heat of sorption is linear rather than logarithmic, as implied in the Freundlich equation. The adsorption experiment data were analysed by Temkin isotherm model in the linearised form,

$$q_e = B \ln C_e + B \ln A$$

where $B = RT/b$, b is the Temkin constant related to heat of sorption (J/mol), A is the equilibrium binding constant corresponding to the maximum binding energy (L/g), R is the gas constant (8.314 J/mol K), and T is the absolute temperature (K). The plot of q_e against $\ln C_e$ is given below.

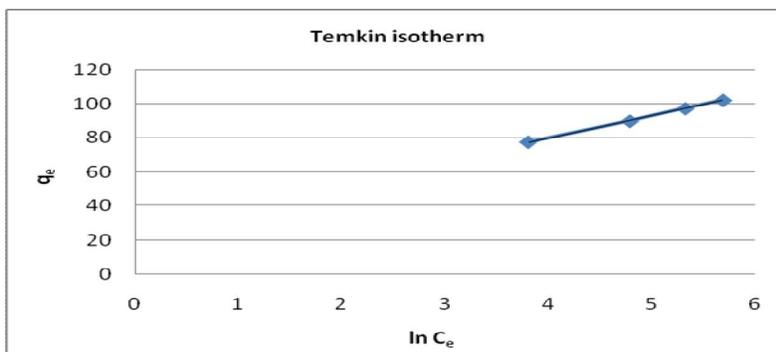


Fig. 8: Temkin Isotherm plot for the adsorption of Copper by Chitosan-Cellulose beads

3.6.4 Table of estimated values of constants of isotherms from the experimental data

The estimated values of the constants of the Isotherm models for the adsorption of Copper by the Chitosan – Cellulose beads have been given in the table below.

Table 1: Table of estimated values of constants of Isotherms

Langmuir Isotherm			Freundlich Isotherm			Temkin Isotherm		
R ²	q _{max} (mg/g)	b	R ²	K _F (mg/g)	1/n	R ²	b (J/mol)	A (L/g)
0.995	111.11	0.0918	0.999	43.95	0.149	0.998	18.96	7.47

Based on the linear regression values ($R^2 > 0.99$) which are considered as a measure of the goodness-of-fit of data, the experimental data followed the order, Freundlich > Temkin > Langmuir.

3.6.5 DUBININ-RADUSHKEVICH ISOTHERM

The Dubinin-Radushkevich isotherm equation is generally used to distinguish between physical and chemical adsorption. It is given in the linearised form as,

$$\ln q_e = K_{DR} \epsilon^2 + \ln q_{max}$$

where q_e is the equilibrium adsorption capacity of the adsorbent (mg/g), q_{max} is the maximum adsorption capacity (mg/g), K_{DR} is the Dubinin-Radushkevich constant (mol^2/kJ^2) and ϵ is Polanyi potential given by,

$$\epsilon = RT \ln (1 + 1/C_e)$$

where R is the gas constant (8.314×10^{-3} kJ/molK), T is the temperature in Kelvin and C_e is the equilibrium concentration of metal ions (ppm). Thus the plot of $\ln q_e$ against ϵ^2 gives a straight line with a slope of K_{DR} and an intercept of q_{max} . The Dubinin-Radushkevich isotherm also gives the mean energy of adsorption by the equation,

$$E = (-2 K_{DR})^{-1/2}$$

If the E value is less than 8 kJ/mol, the process follows physical adsorption, and if the E value lies between 8 and 16 kJ/mol, the process follows chemical adsorption.

The Dubinin-Radushkevich isotherm plot for the experimental data as follows.

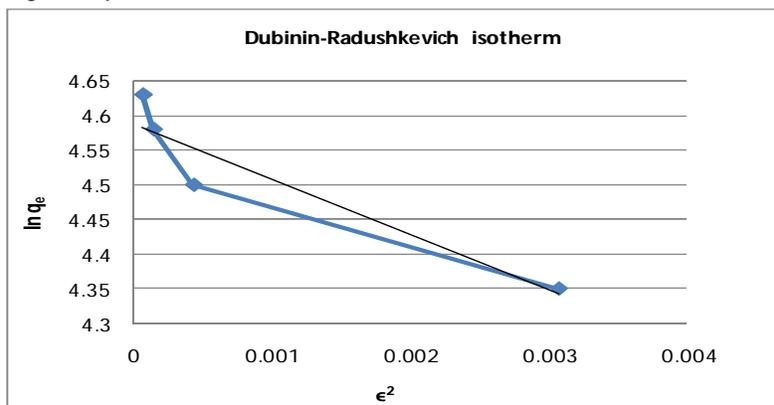


Fig. 9: Dubinin-Radushkevich isotherm plot for the adsorption of Copper by Chitosan-Cellulose beads

From the Dubinin-Radushkevich isotherm plot, the linear regression value R^2 was 0.884. The mean energy of adsorption was found to be 0.079 kJ/mol which was less than 8 kJ/mol, and hence it clearly showed that the adsorption of Copper ions by Chitosan - Cellulose beads was physical adsorption.

3.7 ADSORPTION KINETICS

In order to investigate the mechanism of adsorption and its potential rate controlling steps, kinetic models have been used. The adsorption kinetics of heavy metal ions are

analysed by the pseudo first order, pseudo second order and simple Elovich kinetic models.

3.7.1 PSEUDO FIRST ORDER MODEL

Lagergren's first order rate equation has been most widely used for the adsorption of an adsorbate from an aqueous solution. It is represented as,

$$\ln (q_e - q_t) = \ln q_e - K t$$

where q_e is the equilibrium adsorption capacity (mg/g), q_t is the mass of metal ions adsorbed at time t (mg/g), K is the first order rate

constant (min^{-1}). The pseudo first order considers the rate of occupation of adsorption sites is directly proportional to the number of unoccupied sites. A plot of $\ln(q_e - q_t)$ against t should give a linear relationship for the applicability of the first order kinetic. The

following figure represents the Pseudo First order sorption kinetics of Copper (II) ions on to Chitosan – Cellulose beads for various initial concentrations (100, 200, 300, 400 and 500 ppm) of volume 100 mL (each), adsorbent dose 0.2 g, temperature 28° C and pH 5.2.

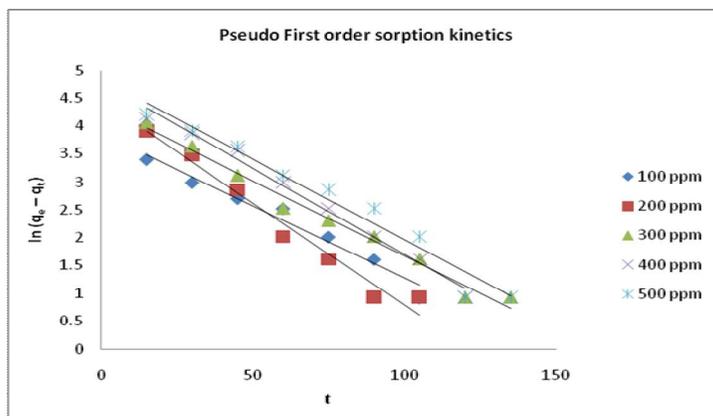


Fig. 10: The Pseudo First order kinetic sorption kinetics for various initial concentrations

3.7.2 PSEUDO SECOND ORDER MODEL

The Pseudo Second order model considers that the rate of adsorption metal ions is based on the square of number of vacant sites on the adsorbent. The pseudo second order rate equation is represented as,

$$t/q_t = 1/(K q_e^2) + t/q_e$$

A plot of t/q_t versus t should give a linear relationship for the applicability of the second

order kinetic. The following figure represents the Pseudo Second order sorption kinetics of Copper (II) ions on to Chitosan – Cellulose beads for various initial concentrations (100, 200, 300, 400 and 500 ppm) of volume 100 mL (each), adsorbent dose 0.2 g, temperature 28° C and pH 5.2.

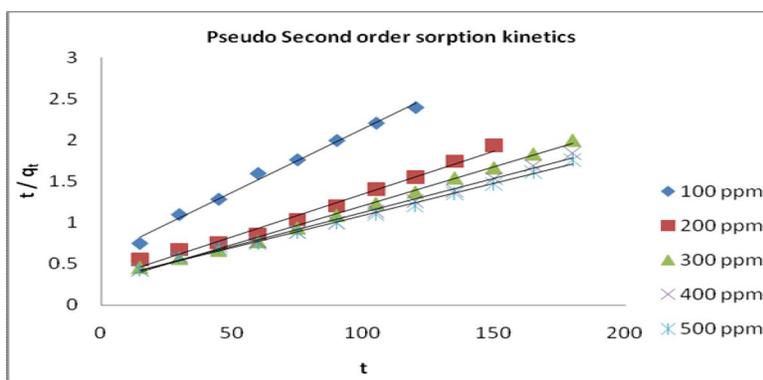


Fig.11: The Pseudo Second order sorption kinetics for various initial concentrations

3.7.3 SIMPLE ELOVICH MODEL

The simple Elovich model is expressed in the form,

$$q_t = \alpha + \beta \ln t$$

where q_t is the amount adsorbed at time t , α and β are the constants obtained from the experiment. A plot of q_t against $\ln t$ should give a linear relationship for the applicability of the simple Elovich kinetic.

The following figure shows the simple Elovich kinetics of Copper (II) ions on to Chitosan – Cellulose beads for various initial concentrations (100, 200, 300 400 and 500 ppm) of volume 100 mL (each), adsorbent dose 0.2 g, temperature 28° C and pH 5.2.

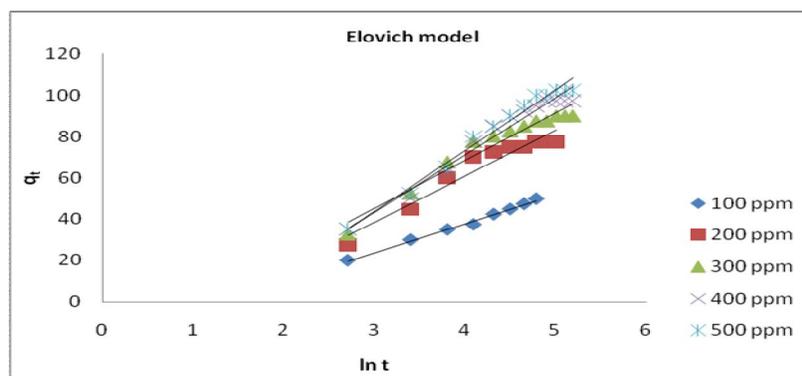


Fig. 12: The simple Elovich sorption kinetics for various initial concentrations

3.7.4 Table of Estimated Parameters of Kinetic Models

The parameters of First order, Second order and Elovich kinetic models are estimated and given below.

Table 2: Table of estimated parameters of Kinetic models

Conc. of aqueous solution (ppm)	First order kinetic model			Second order kinetic model			Elovich kinetic model		
	R ²	q _e (mg/g)	K _{ad} (min ⁻¹)	R ²	q _e (mg/g)	K _{ad} (g/mg min)	R ²	α	β
100	0.970	47.94	0.026	0.993	62.50	4.27*10 ⁻⁴	0.993	-18.69	14.16
200	0.972	84.47	0.036	0.987	90.91	3.94*10 ⁻⁴	0.935	-28.40	23.22
300	0.984	70.59	0.026	0.996	100.0	3.89*10 ⁻⁴	0.946	-23.58	22.96
400	0.989	120.30	0.030	0.992	111.11	2.71*10 ⁻⁴	0.963	-38.97	27.53
500	0.964	140.05	0.028	0.996	125.0	2.08*10 ⁻⁴	0.970	-43.57	29.20

The values of equilibrium adsorption capacity (q_e) obtained from the experiments were 50, 77.5, 90, 97.5 and 102.55 (in mg/g) for the concentrations of aqueous solutions 100, 200, 300, 400 and 500 (in ppm) respectively. Three kinetic models were applied on the experimental data to investigate the suitability. The linear regression values obtained from the Elovich kinetic model were much lower than the other two models and it showed the inapplicability of the model. The linear regression values of Pseudo First order sorption kinetics were also low and the q_e values acquired from the Pseudo First order sorption kinetics were contrasted with the experimental values. But in the case of Pseudo Second order model, the linear regression values were considerably higher (R² > 0.99), and also the calculated q_e values agreed better with the experimental data. Hence it was very clear that the adsorption of Copper by Chitosan - Cellulose beads has followed the Pseudo Second order kinetic model.

3.7.5 INTRA-PARTICLE DIFFUSION MODEL

The adsorption process on a porous adsorbent is generally a multi-step process. In order to analyse the mechanism of the adsorption of Copper by Chitosan-Cellulose beads, the experimental data were tested against the intra-particle diffusion model. The adsorption mechanism of the adsorbate on to the adsorbent follows three consecutive steps: mass transfer across the external film of liquid surrounding the particle, adsorption at the surface of pores and the intra-particle diffusion. The slowest of these steps determines the overall rate of the process. The possibility of intra-particle diffusion resistance which could affect the adsorption is explored by using the intra-particle diffusion model given in the equation,

$$q_t = K t^{1/2} + I$$

where K is the intra-particle diffusion rate constant and I is the intercept. A plot of q_t against t^{1/2} is drawn to analyse the possibility of intra-particle diffusion as the rate determining step. A two stage adsorption mechanism with first rapid and second slower has been observed from the experimental data. The plot

of q_t against $t^{1/2}$ is multi-linear and deviating from the origin, indicating more than one

process has affected the adsorption¹⁸.

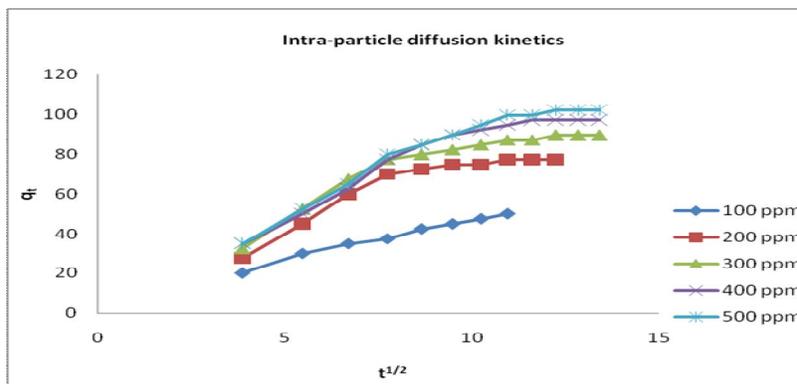


Fig. 13: The intra-particle diffusion kinetics plot for various initial concentrations

3.7.6 THE ESTIMATED PARAMETERS OF INTRA-PARTICLE DIFFUSION MODEL

The parameters of intra-particle diffusion model are estimated and given in the following table.

Table 3: Table of estimated parameters of Intra-Particle diffusion model

Concentration of aqueous solution (ppm)	R ²	K _{ad} (g/mg min ^{0.5})	Intercept (l)
100	0.985	4.076	6.26
200	0.839	5.647	16.60
300	0.844	5.421	25.69
400	0.895	6.631	18.85
500	0.922	7.079	17.34

3.7.7 THE BOYD MODEL

Due to the double nature of intra-particle diffusion (both film and pore diffusion) and in order to determine the actual rate controlling step involved in the sorption process, the kinetic data have been analysed using the model given by Boyd et al¹⁹.

$$F = 1 - (6/\pi^2) \sum_{m=1}^{\infty} [(1/m^2) \exp(-m^2 Bt)]$$

Where F is the fractional attainment of equilibrium at time t and is obtained from the expression:

$$F = q_t / q_e$$

where q_t (mg/g) is the amount of adsorbate taken up at time t and q_e (mg/g) is the maximum equilibrium uptake and

$$B = D_i \pi^2 / r^2$$

where B is the time constant (min^{-1}), D_i is the effective diffusion coefficient of the metal ions in the sorbate phase (cm^2/min), r is the radius of the adsorbent particle (cm), assumed to be spherical, and m is an integer that defines the infinite series solution. Bt is given by the equation:

$$Bt = -0.4977 - \ln(1 - F)$$

Thus the value of Bt can be computed for each value of F , and then plotted against time to configure the so-called Boyd plots²⁰. A straight line passing through origin is indicative of sorption processes governed by particle diffusion mechanism; otherwise they are governed by film diffusion²¹.

The plots of Bt against t for the experimental data of various concentrations have been shown below.

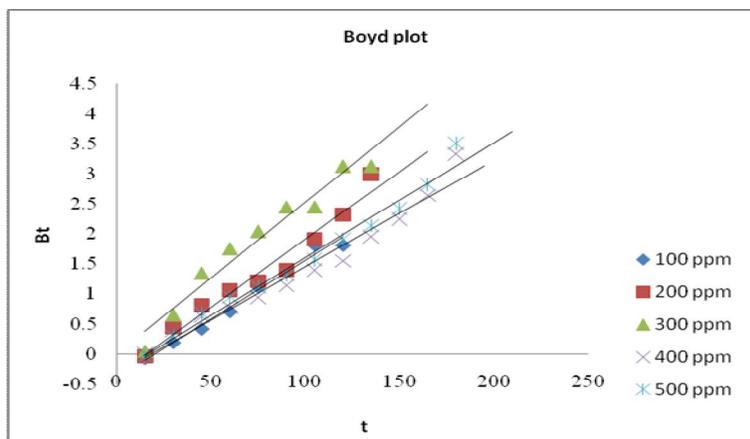


Fig. 14: The Boyd plot (showing markers and trend lines) for the sorption kinetics

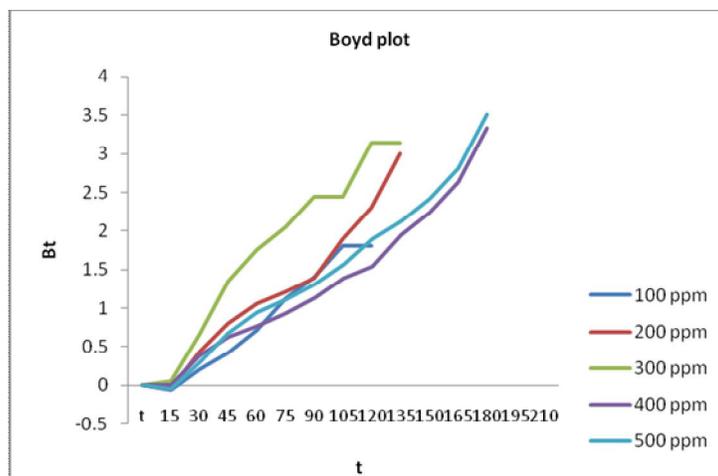


Fig. 15: The Boyd plot (lines graph to show linearity) for the sorption kinetics

The plots very slightly deviated from the origin but they were much linear which revealed that the intra-particle diffusion was the actual rate controlling step of the adsorption process of Copper (II) ions by Chitosan-Cellulose beads.

4. CONCLUSION

Chitosan – Cellulose beads can be successfully used as an adsorbent for Cu^{+2} ions from aqueous solutions. The addition of Cellulose made the adsorbent beads mechanically stronger and considerably denser. The cross-linking with Formaldehyde extended the application of the adsorbent to a wide pH range by enhanced its chemical stability. The maximum adsorption was at a pH of 5. The removal of copper ions increased with agitating time and saturated in about 180 minutes. There was a two stage adsorption mechanism in which the first was rapid and the second was slower has been observed. The adsorption data were best fitted with

Freundlich isotherm and showed multilayer adsorption. The value of mean energy of adsorption (E) from the Dubinin – Radushkevich isotherm has showed that the adsorption of Copper by Chitosan – Cellulose beads followed the physical adsorption. Adsorption kinetics followed the Pseudo Second order kinetic model. The intra-particle diffusion model was used to analyse the sorption mechanism. Boyd plot confirmed that the intra-particle diffusion was the rate controlling step. The results obtained would be useful for the effective application of Chitosan - Cellulose beads as an adsorbent to treat industrial effluents.

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